

# SEASONAL FLUXES OF CO<sub>2</sub>, CH<sub>4</sub> AND N<sub>2</sub>O GREENHOUSE GASES IN VARIOUS MANGROVE SPECIES ON THE COAST OF WEST MUNA REGENCY, SOUTHEAST SULAWESI, INDONESIA

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# Abstract

Seasonal fluxes of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O greenhouse gases in each mangrove species on the coast of West Muna Regency -Southeast Sulawesi, Indonesia conducted in 2019. The largest CO<sub>2</sub> gas flux occurred in April (rainy season) in *B. cylindrica* species with a flux value of 79.29 mg m<sup>-2</sup> h<sup>-1</sup>, while the lowest CO<sub>2</sub> gas flux occurred in July (the dry season) in *S. alba* species with a flux of 6.32 mg m<sup>-2</sup> h<sup>-1</sup>. The largest CH<sub>4</sub> gas flux occurred in April (rainy season) in *R. mucronata* species at 57.16 mg m<sup>-2</sup> h<sup>-1</sup>, while the lowest occurred in July (dry season) in species *B. gymnorrhiza* at 11.38 mg m<sup>-2</sup> h<sup>-1</sup>. The largest N<sub>2</sub>O gas flux occurred in July (dry season) in *S. alba* species at 6.08 mg m<sup>-2</sup> h<sup>-1</sup> and the lowest occurred in July (dry season) in *B. cylindrica* species at 0.43 mg m<sup>-2</sup> h<sup>-1</sup>. Rain duration has a high correlation with CO<sub>2</sub> gas flux in *B. cylindrica* and *B. gymnorrhiza* species and on CH<sub>4</sub> gas flux in *R. mucronata* and *S. alba* species with correlation values of r = 0.6318, r = 5071, r = 0,6371 and r = 0.5076. Rain duration does not correlate with N<sub>2</sub>O gas flux in *S. alba* and *B. gymnorrhiza* species with correlation values of r = 0,0002 and r = 0.0003, respectively.

Key words: Greenhouse gases, Mangrove species, Seasonal fluxes.

### Introduction

Global warming is one of the natural phenomena that can occur because it is triggered by an increase in greenhouse gases in the atmosphere. Global warming or commonly known as climate change can have an impact on aspects of fisheries (Badjeck *et al.*, 2010; Shawket *et al.*, 2019). Some changes that arise as a result of climate change such as rising water temperatures, increased deposition, pH, salinity, oxygen, wind speed, waves and sea level rise that significantly affect ecological conditions or decreased ecosystem services in the sea and freshwater (Cheung *et al.*, 2009; Brander, 2010; Drinkwater *et al.*, 2009; Jones, 2013; Wang *et al.*, 2016) and had an impact on the decline in fisheries production and community livelihoods (Coulthard, 2008; Belhabib *et al.*, 2016; Ohwayo *et al.*, 2016; Rosegrant *et al.*, 2016;

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Valmonte-Santos *et al.*, 2016; Asch *et al.*, 2017; Shaffril *et al.*, 2017a) and can affect food security based on the level of food security vulnerability to climate change (Ding *et al.*, 2017), so social adaptation and policy processes must be adapted both spatially and temporally (de Salamanca *et al.*, 2017; Shaffril *et al.*, 2017b; Singh *et al.*, 2017).

The biggest contributors to global warming were carbon dioxide  $(CO_2)$  and methane  $(CH_4)$  which are produced by various human activities such as burning fossil fuels, motor vehicles and industrial machinery that causes carbon gas to accumulate in the atmosphere (IPCC, 2001; Lang *et al.*, 2011; Oertel *et al.*, 2016). There are three main greenhouse gases identified in the atmosphere, namely  $CO_2$ ,  $CH_4$  and  $N_2O$  (Turunen *et al.*, 2001); Houghton, 2003; Archard *et al.*, 2004; Adi *et al.*, 2009). According to the IPCC, (2007), the percentage

increase in greenhouse gas emissions between 1970 and 2004 has reached 70%.

Mangrove ecosystems are unique ecosystems with puddles that cause high substrate salinity, as well as flooding by tides. Mangrove ecosystems have a high content of organic matter derived from leaves, flowers, branches, twigs and some other tree parts commonly called litter. Besides the litter from mangroves, the input of organic matter from the outside as a result of pressure from anthropogenic activity is also a material to be processed by various decomposer organisms (Giani et al., 1996; Tam and Wong, 2000; Trott and Alongi, 2000; Chang and Yang, 2003; Lovelock et al., 2004). Decomposition of litter and organic matter and various other reactions such as methanogenic, nitrification and denitrification that occur in the mangrove substrate produces the main gases in increasing greenhouse gas emissions, in this case, are CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O (Mosier, 1994; Purvaja and Ramesh, 2001; Hincapie et al., 2002; Kreuswiezer et al., 2003; Alongi et al., 2005; Wang et

*al.*, 2009; Chen *et al.*, 2012; Konnerup *et al.* 2014; Nobrega *et al.* 2016).

CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O gases produced by microbial activities that occur in the soil or substrate will mostly be released into the atmosphere either diffused through the soil or fluxed by plants so that it will increase the concentration of greenhouse gases that affect increasing the temperature of the earth's surface (De Wilde and De Bie, 2000; Allen et al., 2007; Biswas et al., 2007; Davidson, 2009; Dunne et al., 2013; Harley et al., 2015; Castillo et al., 2017; Tullberg et al., 2018) includes fluxes that occur in grasslands (Li et al., 2015). The rate of carbon gas flux (CO<sub>2</sub> and CH<sub>4</sub>) originating from mangrove litter is generally influenced by litter type, litter degradation rate, salinity, temperature, tides, pH and freshwater input (Purvaja and Ramesh, 2001; Kreuzwieser et al., 2003; Arnold et al., 2005; Barnes et al., 2006; Rusmana, 2006; Kristensen et al., 2008a,b; Kone and Borges, 2008; Dutta et al., 2013; Rahman et al., 2018). Besides that, bulk soil density is also a factor influencing the formation of



Fig. 1: Geographical location of the mangrove in the present study. I: Maginti mangrove; II: Central Tiworo mangrove; III: Archipelago Tiworo mangrove; IV: Sawerigadi mangrove.

methane gas in wetlands including mangrove ecosystems (Giani *et al.*, 1996; Chang and Yang, 2003).

The study related to greenhouse gas flux in various types of ecosystems have been carried out, for example, Ye et al., (2000); Kreuzwieser et al., 2003; Arnold et al., (2005); Rusmana, 2006; Chen et al., (2010); Konnerup et al., (2014); Chauhan et al., (2015); Nobrega et al. (2016); Cabezas et al., (2017) and Rahman et al., (2018) in mangrove ecosystems, Midderlburg et al., (2002); Harley et al., (2015); Nirmal-Rajkumar et al., (2008); and Upstill-Goddard and Barnes, (2016) respectively in estuarine waters and Ostrovsky, (2003); Ferron et al., (2007); and Osudar et al., (2015) respectively on the lake, bay and sea waters. However, a study related to greenhouse gas flux to the type of litter from some mangrove species is still very least, especially if it is related to the temperature, salinity, frequency and duration of rain which is a representation of the season.

Indonesia in general consists of two seasons, the dry season and the rainy season. The rainy season occurs in November - May while the dry season occurs in June -October. Information related to greenhouse gas fluctuations from litter of various mangrove species in the dry and rainy seasons, especially in Indonesia, has not yet been found. Whereas the frequency and duration of rain that occurs can influence changes in temperature and salinity, causing fluxes of greenhouse gases to fluctuate. Therefore it is necessary to research to obtain information related to it.

One of the mangrove ecosystem habitats is on the coast of West Muna Regency - Southeast Sulawesi. Mangrove species in this region consist of Bruguiera cylindrica, Bruguiera gymnorrhiza, Rhizophora apiculata, Rhizophora mucronata, Rhizophora stylosa, Sonneratia alba and other species (Rahman *et al.*, 2014) that are still classified as natural because their waters have not been contaminated by industrial waste or domestic waste overload. This is because the majority of the community on the coastal area of West Muna work as farmers and only a small proportion work as fishermen or cultivators. Ecosystem natural conditions supported by low anthropogenic activity stresses and dominant mangrove species are representative enough to illustrate seasonal analysis of greenhouse gas fluxes ( $CO_2$ ,  $CH_4$  and  $N_2O$ ) in various mangrove species in other natural ecosystems in Indonesia.

# **Material and Methods**

### **Description of study area**

This research was conducted in January - December, 2019 as a representation of the rainy season and the dry season. The study sites are in four areas which are mangrove ecosystem habitats. These areas are Maginti District (Station I), Central Tiworo District (Station II), Tiworo Kepulauan District (Station III) and Sawerigadi District (Station IV), West Muna Regency (Fig. 1).

Rahman *et al.*, (2014) reported that mangrove species in this region were dominated by *Bruguiera cylindrica*, *Bruguiera gymnorrhiza*, *Rhizophora apiculata*, *Rhizophora mucronata*, *Rhizophora stylosa* and *Sonneratia alba* with a total densities ranging from 760-1000 trees.ha<sup>-1</sup> and Rahman *et al.*, 2019) mangroves with density values are included in the low category. These species live in different habitat types. *Bruguiera cylindrica* and *Bruguiera gymnorrhiza* live in mudsubstrate habitats, *Rhizophora apiculata*, *Rhizophora mucronata* and *Rhizophora stylosa* live in sandy-mud-

> habitats, while *Sonneratia alba* lives in mud-sand substrate habitats (Bengen, 2004; Noor *et al.*, 2006; Hogarth, 2007; Rahman *et al.*, 2014; Martuti *et al.*, 2019).

> Furthermore, Rahman *et al.*, (2014) states that each mangrove ecosystem has a relatively similar utilization pattern, namely as a fishing ground, timber, ecotourism and around 250 ha of traditional aquaculture ponds, especially in the coastal areas of Maginti District. The conversion of land as ponds, docks, boat moorings and settlements has triggered a decrease in the area and density of mangroves in Indonesia (Ilman *et al.*, 2016) including on the coast of West Muna (Rahman *et al.*, 2019).





### Greenhouse gases samplings and flux measurements

Gas sampling begins by first storing the litter of each mangrove species. Litter collection is done by placing litter traps measuring  $2 \times 4$  m<sup>2</sup> under the canopy of each mangrove species. Next take each of the 600 grams of the wet weight of the litter to be placed and allowed to stand on a square plate  $(1 \times 1 \times 1 \text{ m}^3)$  for 30 days when the litter begins to decay and decompose. When litter begins to decompose, a chamber  $(0.5 \times 0.5 \times 1 \text{ m}^3)$  was placed in a square plate to take gas that results from litter decomposition (Mosier, 1994; Purvaja and Ramesh, 2001; Ye et al., 2000; Hincapie et al., 2002; Kreuswiezer et al., 2003; Alongi et al., 2005; Wang et al., 2009; Chen et al., 2012; Konnerup et al., 2014; Nobrega et al., 2016). Gas samples were taken for 8 hours at intervals of 2 hours (08.00, 10.00, 12.00, 14.00, 16.00). Gas was taken through a syringe tube and put into a 10 ml glass bottle and analyzed using gas chromatography (GC) methods. The litter collection and gas sampling can be presented in fig. 2.

### Measurement of rainy day and daily duration

Measurement of rain frequency and duration was manually carried out by recording its amount and duration in a month using a stopwatch. This is aimed to categorize the summer and the rainy season, so that it can be the analysis of the correlation between the season to greenhouse gas flux of each mangrove species found on the coast of West Muna.

### Data analysis

### • Greenhouse gases fluxes:

Greenhouse gas fluxes (CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O) are analyzed based on differences in mangrove species and seasons represented by the frequency and duration of rainfall. The flux analysis was calculated with the

following formula (IAEA, 1992):

$$F = \frac{dc}{dt} \times \frac{Vch}{Ach} \times \frac{mW}{mV} \times \frac{273,2}{273,2+T}$$

Notes: F = flux (mg m<sup>-2</sup> h<sup>-1</sup>), = change in gas concentration per unit of time (ppm hr<sup>-1</sup>), Vch = volume of chamber (m<sup>3</sup>), Ach = area of chamber (m<sup>2</sup>), mW = molecular weight (g.mole<sup>-1</sup>), mV = mole volume constant (22.4 L), T = average temperature during gas sampling (p C), 273.2 = Kelvin temperature constant.

### Seasonal analysis

In general, the difference between the rainy season and the summer season can be represented by the frequency and rain duration that occurred each month. The season was analysis descriptively-quantitative.

# • Seasonal relationship to greenhouse gases fluxes

The relationship between season and greenhouse gas flux was analyzed using the regression method. This is aimed to see the level of season correlation with the flux of greenhouse gases in each mangrove species. The analysis of correlation level of the season and greenhouse gas flux refers to the classification according to Sarwono, (2009) as in table 1.

### **Result and Discussion**

### Greenhouse gases fluxes

Appendix 1. shows that the greatest CO<sub>2</sub> gas flux is found in *B. cylindrica*, *B. gymnorrhiza*, *R. mucronata* and *R. apiculata*, which are 38.28 mg m<sup>-2</sup> h<sup>-1</sup>, 29.92 mg m<sup>-2</sup> h<sup>-1</sup>, 29.02 mg m<sup>-2</sup> h<sup>-1</sup> and 25.70 mg m<sup>-2</sup> h<sup>-1</sup>. The lowest CO<sub>2</sub> gas flux is found in *R. stylosa* and *S. alba*, namely 18.93 mg m<sup>-2</sup> h<sup>-1</sup> and 16.44 mg m<sup>-2</sup> h<sup>-1</sup>. CH<sub>4</sub> gas flux the largest are found in *R. mucronata*, *B. cylindrica*, *R. stylosa* and *B. gymnorrhiza* were 37.37 mg m<sup>-2</sup> h<sup>-1</sup>,

Appendix 1: Flux of greenhouse gases in various mangrove spesies.

Flux of greenhouse gases (mg m <sup>-2</sup> h <sup>-1</sup> )																		
Time	B. cylindrica			B. gymnorrhiza			R. apiculata			R.mucronata			R. stylosa			S. alba		
(Month)	CO <sub>2</sub>	CH <sub>4</sub>	N <sub>2</sub> O	CO <sub>2</sub>	CH <sub>4</sub>	N <sub>2</sub> O	CO <sub>2</sub>	CH <sub>4</sub>	$N_2O$	CO <sub>2</sub>	CH <sub>4</sub>	N <sub>2</sub> O	CO <sub>2</sub>	CH <sub>4</sub>	N <sub>2</sub> O	CO <sub>2</sub>	CH <sub>4</sub>	$N_2O$
Jan.	35,01	45,65	4,43	27,72	42,85	2,09	13,30	29,30	5,60	16,46	29,34	5,38	12,50	21,30	3,31	8,58	19,61	3,48
Feb.	47,96	49,33	1,46	42,43	35,06	1,32	34,64	25,71	1,72	31,93	29,71	4,70	24,00	28,64	2,99	20,15	28,05	1,58
March	64,53	40,37	1,34	57,17	37,75	1,26	32,58	31,54	5,22	38,59	40,11	4,23	29,80	28,24	1,99	18,17	31,28	4,25
April	79,29	54,76	0,43	62,90	47,83	5,88	43,36	31,57	1,89	51,88	57,16	4,26	35,51	34,80	1,74	23,76	34,18	2,68
May	53,60	48,42	0,93	41,86	37,86	1,60	27,47	22,71	1,83	37,81	51,66	1,26	32,72	49,41	3,13	42,06	29,40	2,26
June	40,80	37,49	2,49	11,65	24,79	5,07	21,07	26,77	5,04	16,09	41,52	5,17	6,52	33,52	2,48	10,43	28,05	6,08
July	7,00	32,93	2,59	13,14	24,14	4,47	18,67	17,29	5,67	10,71	35,39	4,36	15,15	24,58	5,39	6,32	14,95	2,66
Aug.	15,15	27,90	1,28	9,69	28,73	1,81	7,30	20,70	6,25	8,74	32,58	2,71	18,07	33,58	3,95	25,98	21,81	3,13
Sept.	18,07	27,86	3,19	23,52	11,38	5,99	15,18	16,96	4,96	23,99	23,58	1,91	7,46	20,82	3,51	14,24	19,68	5,84
Oct.	7,46	25,88	3,58	10,22	14,83	4,50	35,01	21,79	3,61	45,83	27,47	3,42	16,55	19,41	5,51	6,22	22,30	3,74
Nov.	40,95	21,26	5,95	27,95	20,94	3,53	27,16	22,22	5,22	28,81	36,65	3,92	12,41	28,57	2,61	8,52	17,01	2,01
Dec.	49,49	32,56	1,85	30,83	16,81	4,34	32,65	27,22	3,79	37,41	43,23	3,53	16,46	35,99	2,45	12,87	19,08	1,38
Average	38,28	37,03	2,46	29,92	28,58	3,49	25,70	24,48	4,23	29,02	37,37	3,74	18,93	29,90	3,25	16,44	23,78	3,26

Table 1: Clasification of correlation level (Sarwono, 2009).

R	Correlation level
0	No correlation
0,00-0,25	Low correlation
0,25-0,50	Moderat
0,50-0,75	High correlation
0,75-0,99	Very high correlation
1	Perfect

37.03 mg m<sup>-2</sup> h<sup>-1</sup>, 29.90 mg m<sup>-2</sup> h<sup>-1</sup> and 28.58 mg m<sup>-2</sup> h<sup>-1</sup>, respectively. The lowest CH<sub>4</sub> gas flux is found in *R. apiculata* namely 24.48 mg m<sup>-2</sup> h<sup>-1</sup> and *S. alba* namely 23.78 mg m<sup>-2</sup> h<sup>-1</sup>. The largest N<sub>2</sub>O gas flux is *R. apiculata*, *R. mucronata* and *B. gymnorrhiza* were 4.23 mg m<sup>-2</sup> h<sup>-1</sup>, 3.74 mg m<sup>-2</sup> h<sup>-1</sup> and 3.49 mg m<sup>-2</sup> h<sup>-1</sup>, respectively. The lowest N<sub>2</sub>O gas flux is found in *B. cylindrica*, *R. stylosa* and *S. alba* were 2.46 mg m<sup>-2</sup> h<sup>-1</sup>, 3.25 mg m<sup>-2</sup> h<sup>-1</sup> and 3.26 mg m<sup>-2</sup> h<sup>-1</sup>, respectively.

### Rainy days and daily duration

In 2019/2020 the rainy season experienced a time shift from April-August and January-June 2019, in several regions in Sulawesi. Based on the analysis of the rain frequency and duration throughout 2019, it was found that West Muna Regency was one of the regions that experienced a shift in season, with rainfall from January to July followed by the dry season which started in August. The peak of the rainfall occurred in April with a frequency of 17 days month<sup>-1</sup> with a duration of 2.03 hours day<sup>-1</sup> and the lowest in September with a frequency of 1 day month<sup>-1</sup> and a duration of 0.01 hours day<sup>-1</sup>.

However, there was low-frequency rainfall in September, November and December, therefore, it was concluded that subsequent months were called "wet-dry season" (Fig. 3).

### Seasonal relationship to greenhouse gases fluxes

The correlation of rain duration to greenhouse gas flux in each species is relatively different. Rain duration



Fig. 3: Rainy days and daily duration in 2019.

has a low correlation with CO<sub>2</sub> gas flux in *S. alba* and a strong correlation with CO<sub>2</sub> gas flux in *B. cylindrica* and *B. gymnorrhiza*. Rain duration has a low correlation with CH<sub>4</sub> gas flux in *R. stylosa*, but has a low correlation with CH<sub>4</sub> gas flux in *R. mucronata* and *S. alba*. While the correlation of rain duration to N<sub>2</sub>O gas flux is low in *B. gymnorrhiza*, *R. apiculata*, *R. mucronata* and *S. alba*, but moderately correlated with N<sub>2</sub>O gas flux in *B. cylindrica* and *R. stylosa* (Table 2).

# CO,

In general, CO, gas fluxes for each mangrove species are relatively different and their fluctuations tend to follow fluctuations in rain duration throughout 2019 (Fig. 4). In January 2019 which is the rainy season, it rains for 4 days with an average duration of 0.17 hour day<sup>-1</sup>. Species with the largest CO<sub>2</sub> gas flux this month are B. cylindrica and *B. gymnorrhiza* with values of 35.01 mg m<sup>-2</sup>  $h^{-1}$ and 27.72 mg m<sup>-2</sup> h<sup>-1</sup>, respectively. While the species with the smallest flux is S. alba with a value of 8.58 mg  $m^{-2} h^{-1}$ . CO<sub>2</sub> gas flux in the *B. cylindrica* species, *B.* gymnorrhiza, R. mucronata, R. apiculata and R. stylosa increased to reach the largest values in April, with values of 79.29 mg m<sup>-2</sup> h<sup>-1</sup>, 62.90 mg m<sup>-2</sup> h<sup>-1</sup>, 51.88 mg m<sup>-2</sup> h<sup>-1</sup>, 43.36 mg m<sup>-2</sup> h<sup>-1</sup> and 35.51 mg m<sup>-2</sup> h<sup>-1</sup>, respectively. The flux is in line with the frequency and duration of rain, namely 17 days month<sup>-1</sup> and 2.03 hour day<sup>-1</sup>, which is the peak of the rainy season. Entering the dry season (June-October) the frequency and duration of raindrops dramatically, especially in August and October where rain does not occur at all. This causes the CO<sub>2</sub> gas flux to also experience a significant decrease in each mangrove species. The lowest CO, gas flux for R. apiculata, R. mucronata and B. gymnorrhiza species occurred in August with values of 7.30 mg m<sup>-2</sup> h<sup>-1</sup>, 8.74 mg m<sup>-2</sup> h<sup>-1</sup> and 9.69 mg m<sup>-2</sup> h<sup>-1</sup>, respectively. While the



Fig. 4: Seasonal fluxes of CO<sub>2</sub> gases in various mangrove spesies on the coast of West Muna Regency. Bc =Bruguiera cylindrica, Bg = Bruguiera gymnorrhiza,Ra = Rhizophora apiculata, Rhizophora mucronata,Rhizophora stylosa, Rs = Sonneratia alba.

Green	Dalla	Species												
house	Daily	B.cylii	ndrica	B.gymnorrhiza		R.apiculata		R.mu	cronata	R. stylosa		<i>S. a</i>	lba	
gases	duration	Flux	r	Flux	r	Flux	r	Flux	r	Flux	r	Flux	r	
CO <sub>2</sub>	0,00	35,01	0,6318	27,72	0,5071	13,30	0,1426	16,46	0,3051	12,50	0,3356	8,58	0,0925	
	0,00	47,96	a	42,43	а	34,64	c	31,93	b	24,00	b	20,15	с	
	0,01	64,53		57,17		32,58		38,59		29,80		18,17		
	0,08	79,29		62,90		43,36		51,88		35,51		23,76		
	0,17	53,60		41,86		27,47		37,81		32,72		42,06		
	0,26	40,80		11,65		21,07		16,09		6,52		10,43		
	0,34	7,00		13,14		18,67		10,71		15,15		6,32		
	0,71	15,15		9,69		7,30		8,74		18,07		25,98		
	0,98	18,07		23,52		15,18		23,99		7,46		14,24		
	1,22	7,46		10,22		35,01		45,83		16,55		6,22		
	1,33	40,95		27,95		27,16		28,81		12,41		8,52		
	2,03	49,49		30,83		32,65		37,41		16,46		12,87		
CH <sub>4</sub>	0,00	45,65	0,4788	42,85	0,3609	29,30	0,4253	29,34	0,6371	21,30	0,2244	19,61	0,5076	
	0,00	49,33	b	35,06	b	25,71	b	29,71	a	28,64	с	28,05	а	
	0,01	40,37		37,75		31,54		40,11		28,24		31,28		
	0,08	54,76		47,83		31,57		57,16		34,80		34,18		
	0,17	48,42		37,86		22,71		51,66		49,41		29,40		
	0,26	37,49		24,79		26,77		41,52		33,52		28,05		
	0,34	32,93		24,14		17,29		35,39		24,58		14,95		
	0,71	27,90		28,73		20,70		32,58		33,58		21,81		
	0,98	27,86		11,38		16,96		23,58		20,82		19,68		
	1,22	25,88		14,83		21,79		27,47		19,41		22,30		
	1,33	21,26		20,94		22,22		36,65		28,57		17,01		
	2,03	32,56		16,81		27,22		43,23		35,99		19,08		
N <sub>2</sub> O	0,00	4,43	0,2820	2,09	0,0003	5,60	0,2120	5,38	0,1075	3,31	0,3781	3,48	0,0002	
	0,00	1,46	b	1,32	с	1,72	с	4,70	с	2,99	b	1,58	c	
	0,01	1,34		1,26		5,22		4,23		1,99		4,25		
	0,08	0,43		5,88		1,89		4,26		1,74		2,68		
	0,17	0,93		1,60		1,83		1,26		3,13		2,26		
	0,26	2,49		5,07		5,04		5,17		2,48		6,08		
	0,34	2,59		4,47		5,67		4,36		5,39		2,66		
	0,71	1,28		1,81		6,25		2,71		3,95		3,13		
	0,98	3,19		5,99		4,96		1,91		3,51		5,84		
	1,22	3,58		4,50		3,61		3,42		5,51		3,74		
	1,33	5,95		3,53		5,22		3,92		2,61		2,01		
	2,03	1,85		4,34		3,79		3,53		2,45		1,38		
		a: hig	h correlati	on, b: moo	derat, c: lo	w correla	tion, unit o	of greenho	ouse gases	= mg m <sup>-</sup>	$^{2} h^{-1}$			

Tabel 2: Seasonal relationship to greenhouse gases fluxes.

lowest flux for *R. stylosa, B. cylindrica* and *S. alba* species respectively occurred in June, July and October with flux values of 6.52 mg m<sup>-2</sup> h<sup>-1</sup>, 7.00 mg m<sup>-2</sup> h<sup>-1</sup> and 6.23 mg m<sup>-2</sup> h<sup>-1</sup>, respectively. CO<sub>2</sub> gas fluxes in each species then increase again when entering the rainy season, especially in December 2019 with a value of 49.49 mg m<sup>-2</sup> h<sup>-1</sup> for *B. cylindrica*, 30.83 mg m<sup>-2</sup> h<sup>-1</sup> for *B. gymnorrhiza*, 32.65 mg m<sup>-2</sup> h<sup>-1</sup> for *R. apiculata*, 37.41 mg m<sup>-2</sup> h<sup>-1</sup> for *R. mucronata*, 16.46 mg m<sup>-2</sup> h<sup>-1</sup> for *R. stylosa* and 12.87 mg m<sup>-2</sup> h<sup>-1</sup> for *S. alba*. The value of CO<sub>2</sub> gas flux is relatively low when compared to the

report of Arnold *et al.*, (2005); Chen *et al.*, (2010, 2012) and Rahman *et al.*, (2018).

The difference in CO<sub>2</sub> gas flux in each species is in line with the value of its correlation with the seasons. Seasonal correlation to CO<sub>2</sub> gas flux in *B. cylindrica* and *B. gymnorrhiza* species is relatively strong. The correlation value of these species was r = 0.6318 for *B. cylindrica* and r = 5071 for *B. gymnorrhiza*. Seasonal correlation to CO<sub>2</sub> gas flux is low in species *R. apiculata* (r=0.1426) and *S. alba* (r=0.0925) and moderate in species

### *R. mucronata* (= 0.3061) and *R. stylosa* (r = 0.3051).

The occurrence of differences in gas flux in the rainy season and the dry season is caused by temperature and salinity fluctuations caused by freshwater input (Kreuzwieser et al., 2003; Barnes et al., 2006; Kristensen et al., 2008ab; Kim, 2015), especially during heavy rain. In the rainy season the temperature and salinity of the waters decreases making it possible for microorganisms to carry out microbiological processes especially in litter decomposition (Yunasfi, 2006; Alongi et al., 2000, 2001; 2004; Alongi, 2009; Afdal et al., 2012) so that gas flux the resulting yield also increases, on the contrary, the dry season temperature and salinity of the waters increase which causes pressure for microorganisms so that they cannot carry out the decomposition process optimally. Chen *et al.*, (2010) state that the production of greenhouse gases is significantly influenced by microbiological processes.

Furthermore, differences in CO<sub>2</sub> gas flux in each species are also caused by zoning systems and differences in vegetation. *B. cylindrica* and *B. gymnorrhiza* species have a greater flux than other species because they are in a closed zone with mud habitat, *R. apiculata*, *R. mucronata* and *R. stylosa* species are in the middle zone with sandy-mud habitat, while *S. alba* species are in an open zone with muddy-sand habitat (Bengen, 2004; Noor et al., 2006; Hogarth, 2007; Rahman et al., 2014; Martuti et al., 2019). Substrate differences show differences in soil total carbon, soil bulk density and salinity (Corredor et al., 1999; Purvaja and Ramesh, 2001; Kreuzwieser et al., 2003; Lekphet et al., 2005; Huang et al., 2009).

### CH<sub>4</sub>

 $CH_4$  gas fluctuations are relatively the same as  $CO_2$ 



Fig. 5: Seasonal fluxes of  $CH_4$  gases in various mangrove spesies on the coast of West Muna Regency. Bc =Bruguiera cylindrica, Bg = Bruguiera gymnorrhiza,Ra = Rhizophora apiculata, Rhizophora mucronata,Rhizophora stylosa, Rs = Sonneratia alba.

gas fluctuations. The largest gas flux occurs in the rainy season and the lowest in the dry season. At the peak of the rainy season that occurred in April, the species with the largest gas flux were *R. mucronata*, *B. cylindrica* and *R. stylosa* with gas flux were 57.16 mg m<sup>-2</sup> h<sup>-1</sup>, 54.76 mg m<sup>-2</sup> h<sup>-1</sup> and 49.41 mg m<sup>-2</sup> h<sup>-1</sup>, respectively. In the dry season the lowest gas flux occurs in September in *B. gymnorrhiza*, *R. apiculata* and *R. mucronata* species with the gas flux were 11.38 mg m<sup>-2</sup> h<sup>-1</sup>, 16.96 mg m<sup>-2</sup> h<sup>-1</sup> and 23.58 mg m<sup>-2</sup> h<sup>-1</sup>, respectively. The lowest gas flux also occurred in July in *S. alba* (flux = 14.95 mg m<sup>-2</sup> h<sup>-1</sup>), in October at *R. stylosa* (flux -19.41 mg m<sup>-2</sup> h<sup>-1</sup>) and in November in *B. cylindrica* (flux = 21.26 mg m<sup>-2</sup> h<sup>-1</sup>) (Fig. 5). This shows that the season influences methane gas emissions that occur in mangrove ecosystems.

Table 2 shows that the influence of seasons on CH. gas flux tends to be different for each mangrove species. Season has a strong correlation to flux in species R. mucronata (r = 0.6371) and S. alba (r = 0.5076), moderate correlation to flux in species B. cylindrica (r = 0.4788), B. gymnorrhiza (r = 0.3609) and R. apiculata (r = 0.4253), as well as a low correlation with flux in the *R. apiculata* species (r = 0.2244). Furthermore, the correlation of seasons to CH<sub>4</sub> flux in each species tends to be different when compared to CO<sub>2</sub>. Season is highly correlated to CO<sub>2</sub> gas flux in *B. cylindrica* and *B.* gymnorrhiza species, but moderate to CH<sub>4</sub> gas flux. Season is low correlated to CO<sub>2</sub> gas flux in *R. apiculata* species but moderate to CH<sub>4</sub> gas flux. Season is moderately correlated with CO<sub>2</sub> gas flux in R. apiculata species and low in S. alba species, but high in  $CH_4$  gas flux in the same species. In the same species (*R. stylosa*), the season has a moderate correlation with CO<sub>2</sub> gas flux, but it has a low correlation with CH<sub>4</sub> gas flux. This shows that many factors influence the occurrence of CO<sub>2</sub> and CH<sub>4</sub> gas flux, especially related to microorganisms producing these gases or other factors. The formation of



Fig. 6: pH and salinity in mangrove ecosystem on the coast of West Muna Regency.

methane gas in mangrove ecosystems is strongly influenced by the activity of methanogenic bacteria. Methanogenic bacteria are strongly influenced by bulk soil density and are sensitive to pH in the range of 6-8 (Giani *et al.*, 1996). In this study, waters pH ranged from 6.8-7.6 and was the maximum pH for methane formation in wetlands (Giani *et al.*, 1996; Chang and Yang, 2003). Another factor that triggers methane gas formation in wetlands is salinity. The best salinity to methane formation was from 15-20 ppt (Purvaja and Ramesh, 2001; Arnold *et al.*, 2005; Kone and Borges, 2008; Dutta *et al.*, 2013). In this study, salinity ranged from 28.31-30.11 ppt (Fig. 6) and is a poor salinity in the formation of methane gas (Chen *et al.*, 2010).

In general, CH4 gas flux in each species is lower than the report of Ye *et al.*, (2000); Arnold *et al.*, (2005); Chen *et al.*, (2010); Konnerup *et al.*, (2014); Chauhan *et al.*, (2015); Cabezas *et al.*, (2017) and Rahman *et al.*, (2018) which are also carried out in mangrove ecosystems. Methane gas flux is also lower than that of Midderlburg *et al.*, (2002); Harley *et al.*, (2015); Nirmal-Rajkumar *et al.*, (2008) and Upstill-Goddard and Barnes (2016) in estuarine waters, respectively, Mastepanov *et al.*, (2008) and Kim, (2015) in the tundra, as well as report by Ostrovsky (2003); Ferron *et al.*, (2007); and Osudar *et al.*, (2015) each conducted in a lake, bay and sea waters. **N,O** 

In contrast to CO<sub>2</sub> and CH<sub>4</sub> gas, N<sub>2</sub>O gas flux occurs higher in the dry season compared to the rainy season. In the rainy season, the smallest N<sub>2</sub>O gas flux was found in *B. cylindrica* (flux = 0.43 mg m<sup>-2</sup> h<sup>-1</sup>) in April, *R. apiculata* (flux = 1.25 mg m<sup>-2</sup> h<sup>-1</sup>) in May, *B. gymnorrhiza* (flux = 1.26 mg m<sup>-2</sup> h<sup>-1</sup>) in March, *S. alba* (flux =1.38 mg m<sup>-2</sup> h<sup>-1</sup>) in December, *R. apiculata* (flux=



Fig. 7: Seasonal fluxes of  $CH_4$  gases in various mangrove spesies on the coast of West Muna Regency. Bc =Bruguiera cylindrica, Bg = Bruguiera gymnorrhiza,Ra = Rhizophora apiculata, Rhizophora mucronata,Rhizophora stylosa, Rs = Sonneratia alba.

1.72 mg m<sup>-2</sup> h<sup>-1</sup>) in May and R. stylosa (flux = 1.74 mg  $m^{-2} h^{-1}$ ) in April. In the dry season, the largest gas flux is found in the *R. apiculata* species (flux =  $6.25 \text{ mg m}^{-2} \text{ h}^{-1}$ ) in August, S. alba (6.08 mg m<sup>-2</sup> h<sup>-1</sup>) in June, B. gymnorrhiza (flux = 5.99 mg m<sup>-2</sup> h<sup>-1</sup>) in September, B. cylindrica (flux = 5.95 mg m<sup>-2</sup> h<sup>-1</sup>) in November, R. stylosa (flux = 5.51 mg m<sup>-2</sup> h<sup>-1</sup>) in October and R. *mucronata* (flux = 5.17 mg m<sup>-2</sup> h<sup>-1</sup>) in January (Fig. 7). Thus indicates that the frequency and duration of rain did not have a significant effect on N<sub>2</sub>O gas flux in each mangrove species. The results of the analysis show that rain duration has a low correlation with N<sub>2</sub>O gas flux in S. alba species (r = 0.0002), B. gymnorrhiza (r =0,0003), R. mucronata (r = 0.1075) and R. apiculata (r= 0.2120 ), as well as moderate correlations with B. *cylindrica* (r = 0.3781) and *R. stylosa* (r = 0.2820). The N<sub>2</sub>O flux value is lower than report of Pathak, (1999); Zheng et al., (2000); Dalal et al., (2003); Kreuzwieser et al., (2003); Arnold et al., (2005); Nirmal-Rajkumar et al., (2008); Huang et al., (2014); Konnerup et al., (2014); and Rahman et al., (2018) both in mangrove ecosystems and in other ecosystems and higher when compared with Chen et al., (2010) reports in mangrove ecosystems in South China with a flux range of 5.6-224.14 mg m<sup>-2</sup> h<sup>-1</sup>.

This difference occurs because in this study the type and amount of litter is limited to each mangrove species that has been trapped in a closed manner so that, no input of organic matter from the outside is mixed and is part of the litter decomposition process, consequently the resulting gas flux is purely from 600 g decomposition litter wet weight of each mangrove species. The low greenhouse gas flux in each mangrove species also occurs because the mangrove ecosystem on the coast of West Muna is still natural and has not been contaminated by industrial waste, except by small amounts of domestic waste. Domestic wastes that enter the mangrove ecosystem waters are more susceptible to dilution with the tides, so that there is no long-term waste left in the ecosystem water bodies.

Also, the low N<sub>2</sub>O flux value indicates that temperature and salinity are negatively correlated to N<sub>2</sub>O gas production. Chen *et al.*, (2010) reported that N<sub>2</sub>O production was positively correlated to NH<sub>4</sub><sup>+</sup> and negatively to NO<sub>3</sub><sup>-</sup>. This indicates that N<sub>2</sub>O gas production is dominated by nitrification and denitrification process which is supported by oxygen and inorganic nitrogen content (Corredor *et al.*, 1999; Purvaja and Ramesh, 2001; Kreuswiezer *et al.*, 2003; Rusmana, 2006; van den Heuvel *et al.*, 2009; Chen *et al.*, 2010). According to Huang *et al.*, (2014) the combination of the availability of ammonium (NH<sub>4</sub><sup>+</sup>), nitrate (NO<sub>3</sub><sup>-</sup>) and oxygen is positively correlated to the production of  $N_2O$  gas (r = 0.764).

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# References

- Achard, F., H.D. Eva, P. Mayaux, H.J. Stibig and A. Belward (2004). Improved estimates of net carbon emissions from land cover change in the tropics for the 1990s. *Glob. Biogeochem. Cyc.*, **18**: 1-11.
- Adi, S., F. Salim, T. Suryati, T. Handayani, Hartini and N. Sudiana (2009). Carbon emission of wetland, agriculture and forestry in Indonesia. J. of Environ. Tech., (Special Edition): 1-12.
- Afdal, F. Richardus, Kaswadji and F.K. Alan (2012). CO<sub>2</sub> gases exchange sea-to-air in aquatif of Nasik Peninsula, Belitung. J. Segara., 8: 9-17.
- Allen, D.E., R.C. Dalal, H. Rennenberg, R.L. Meyer, S. Reeves and S. Schmidt (2007). Spatial and temporal variation of nitrous oxide and methane flux between subtropical mangrove sediments and the atmosphere. *Soil. Biol. Biochem.*, **39**: 622-631.
- Allen, D.E., R.C. Dalal, H. Rennenberg and S. Schmidt (2011). Seasonal variation in nitrous oxide and methane emissions from subtropical estuary and coastal mangrove sediments, Australia. *Plant Biol.*, **13**: 126-133.
- Alongi, D.M., F. Tirendi, L.A. Trott and T.T. Xuan (2000). Benthic decomposition rates and pathways in plantations of the mangrove Rhizophora apiculata in the Mekong delta, Vietnam. *Mar. Ecol. Prog. Ser.*, **194:** 87-101.
- Alongi, D.M., G Wattayakorn, J. Pfitzner, F. Tirendi, I. Zagorskis, GJ. Brunskill, A. Davidson and B.F. Clough (2001). Organic carbon accumulation and metabolic pathways in sediments of mangrove forests in southern Thailand. *Mar. Geol.*, **179:** 85-103.
- Alongi, D.M., A. Sasekumar, V.C. Chong, J. Pfitzner, L.A. Trott, F. Tirendi, P. Dixon and G.J. Brunskill (2004). Sediment accumulation and organic material ûux in a managed mangrove ecosystem: estimates of land-ocean-atmosphere exchange in peninsular Malaysia. *Mar. Geol.*, **208**: 383-402.
- Alongi, D.M., J. Pfitzner, L.A. Trott, F. Tirendi and D.W. Klumpp (2005). Rapid sedimentation and microbial mineralization in mangrove forest of the Jiulongjiang estuary, China. *Estuar. Coast. Shelf. Sci.*, 63: 605-618.
- Alongi, D.M. (2009). The energetics of mangrove forests. London, UK: Springer Press.
- Arnold, K.V., P. Weslien, M. Nillson, B.H. Svensson and L. Klemedtsson (2005). Fluxes of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O from drained coniferous forests on organic soils. *For. Ecol. Manage.*, **210**: 239-254.
- Asch, R.G., W.W.L. Cheung and G. Reygondeau (2007). Future marine ecosystem drivers, biodiversity and Fisheries

maximum catch potential in Pacific Island countries and territories under climate change. *Mar. Pol.*, <u>http://dx.doi.org/10.1016/j.marpol.2017.08.015.</u>

- Badjeck, M.C., E.H. Allison, A.S. Halls and D.K. Dulvy (2010). Impacts of climate variability and change on fishery-based livelihoods. *Mar. Pol.*, 34: 375-383.
- Barnes, J., R. Ramesh, R. Purvaja, A.N. Rajkumar, B.S. Kumar, K. Krithika, K. Ravichandran, G. Uher and R. Upstill-Goddard (2006). Tidal dynamics and rainfall control  $N_2O$  and  $CH_4$  emissions from a pristine mangrove creek. *Geophys. Res. Lett.*, **33:** L15405.
- Belhabib, D., V.W.Y. Lam and W.W.L. Cheung (2016). Overview of West African fisheries under climate change: impacts, vulnerability and adaptif responses of the artisanal and industrial sectors. *Mar. Pol.*, **71:** 15-28.
- Bengen, D.G. (2004). Guidelines of introducing and management of mangrove ecosystem. Bogor, ID: CCRMS-IPB University.
- Biswas, H., S.K. Mukhopadhyay, S. Sen and T.K. Jana (2007). Spatial and temporal pattern of methane dynamics in the tropical mangrove dominated estuary, NE coast of Bay of Bengal, India. J. Mar. Syst., 68: 55-64.
- Brander, K. (2010). Impacts of climate change on fisheries. J. Mar. Sys., **79:** 389-402.
- Cabezaz, A., W.J. Mitsch, C. MacDonnel, L. Zhang, F. Bydalek and A. Lasso (2007). Methane emissions from mangrove soils in hydrologically disturbed and reference mangrove tidal creeks in southwest Florida. *Ecol. Engine.*, <u>http:// dx.doi.org/10.1016/j.ecoleng.2017.08.041</u>.
- Castillo, J.A.A., A.A. Apan, T.N. Maraseni and S.G. Salmo III (2017). Soil greenhouse gas fluxes in tropical mangrove forests and in land use on deforested mangrove lands. *Catena.*, **159**: 60-69.
- Chang, T.C., and S.S. Yang (2003). Methane emission from wetland in Taiwan. *Atmos. Environ.*, **37:** 4551-4558.
- Chauhan, R., A. Datta, A.L. Ramanathan and T.K. Adhya (2015). Factors influencing spatio-temporal variation of methane and nitrous oxide emission from a tropical mangrove of eastern coast of India. *Atmos. Environ.*, **107**: 95-106.
- Chen, G.C., N.F.Y. Tam and Y. Ye (2010). Summer fluxes of atmospheric greenhouse gases N<sub>2</sub>O, CH<sub>4</sub> and CO<sub>2</sub> from mangrove soil in South China. *Sci. Tot. Environ.*, **408**: 2761-2767.
- Chen, G.C., N.F.Y. Tam and Y. Ye (2012). Spatial and seasonal variations of atmospheric N<sub>2</sub>O and CO<sub>2</sub> fluxes from a subtropical mangrove swamp and their relationships with soil characteristcs. *Soil. Biol. Biochem.*, **48**: 175-181.
- Cheung, W.W.L., V.W.Y. Lam, J.L. Sarmiento, K. Kearney, R. Watson and D. Pauly (2009). Projecting global marine biodiversity impates under climate change scenario. *Fish Fisher*, DOI 10.1111/j.1467-2979.2008.00315.x.
- Corredor, J.E., J.M. Morell and J. Bauz (1999). Atmospheric nitrous oxide ûuxes from mangrove soils. *Mar. Pollut. Bull.*, 38: 473-478.
- Coulthard, S. (2008). Adapting to environmental change in artisanal fisheries-insight from a South Indian Lagoon. *Glob. Environ. Cha.*, **18**: 479-489.

- Davidson, E.A. (2009). The contribution of manure and fertilizer nitrogen to atmospheric nitrous oxide since 1860. Nat. Geosci., 2: 659-662.
- Dalal, R.C., W.J. Wang, G. Robertson, Philip and W.J. Parton (2003). Nitrous oxide emission from Australian agricultural lands and mitigation options: A review. *Aust. J. Soil. Res.*, 4: 165-195.
- De-Salamanca, Á.E., R. Díaz-Sierra, R.M. Martín-Aranda and M.J. Santos (2017). Environmental impacts of climate change adaptation. *Environ. Imp. Assess. Rev.*, 64: 87-96.
- De-Wilde, H.P.J. and M.J.M. De Bie (2000). Nitrous oxide in the Schelde estuary: production by nitrification and emission to the atmosphere. *Mar. Chem.*, **69**: 203-216.
- Ding, Q., X. Chen, R. Hilborn and Y. Chen (2017). Vulnerability to impacts of climate change on marine fisheries and food security. *Mar. Pol.*, 83: 55-61.
- Drinkwater, K.F., G. Beaugrand, M. Kaeriyama, S. Kid, G. Ottersen, R.I. Perry, H.O. Pörtner, J.J. Polovina and A. Takasuka (2010). On the processes linking climate to ecosystem changes. J. Mar. Sys., 79: 374-388.
- Dunne, J.A., S.C. Jackson and J. Harte (2013). Greenhouse effect. *Encyclo, Biodiv.*, **4:** 277-293.
- Dutta, M.K., C. Chowdhury, T.K. Jana and S.K. Mukhopadhyay (2013). Dynamics and exchange fluxes of methane in the estuarine mangrove environment of the Sundarbans, NE coast of India. *Atmos. Environ.*, **77:** 631-639.
- Ferrón, S., T. Ortega, A. Gómez-Parra and J.M. Forja (2007). Seasonal study of dissolved  $CH_4$  and  $N_2O$  in a shallow tidal system of the Bay of Cádiz (SW Spain). *J. Mar. Syst.*, **66:** 244-57.
- Giani, L., Y. Bashan, G. Holguin and A. Strangmann (1996). Characteristics and methanogenesis of the Balandra lagoon mangrove soils, Baja California Sur, Mexico. *Geoder.*, 72: 149-160.
- Harley, J.F., L. Carvalho, B. Dudley, K.V. Heal, R.M. Rees and U. Skiba (2015). Spatial and seasonal fluxes of the greenhouse gases  $N_2O$ ,  $CO_2$  and  $CH_4$  in a UK macrotidal estuary. *Estuar*. *Coast. Shelf. Sci.*, **153**: 62-73.
- Hincapie, M.M., J.M. Morell and J.E. Corredor (2002). Increase of nitrous oxide flux to the atmosphere upon nitrogen addition to red mangroves sediments. *Mar. Pollut. Bull.*, 44: 992-996.
- Hogarth, P.J. (2007). The Biologis of Mangroves and Seagrasses. America, US. Oxford University Press.
- Houghton, R.A. (2003). Revised estimates of the annual net flux of carbon to the atmosphere from changes in land use and land management 1850-2000. *Tell.*, **55B:** 378-390.
- Huang, B., K. Yu and R.P. Gambrell (2009). Effects of ferric iron reduction and regeneration on nitrous oxide and methane emissions in a rice soil. *Chemos.*, **74:** 481-486.
- Huang, J., Y. Chen, P. Sui, S. Nie, W. Gao (2014). Soil nitrous oxide emissions under maize-legume intercropping system in the North China Plain. J. Integra. Agricult., 13: 1363-1372.
- I.A.E.A. International Atomic Energy Agency (1992). Manual on Measurement of Methane and Nitrous Oxide Emission

from Agricultural. Vienna, AT: IAEA.

- Ilman, M.P., P. Dargusch, P. Dart and Onrizal (2016). A historical analysis of the drivers of loss and degradation of Indonesia's mangroves. *Land Use Pol.*, **54**: 448-459.
- I.P.C.C. (1990). Climate Change. The intergovernmental panel on climate change scientiûc assessment. Cambridge, UK: Cambridge University Press.
- I.P.C.C. (2001). Climate Change. The Intergovernmental oanel on climate change the scientific basis. Cambridge UK: Cambridge University Press.
- I.P.C.C. (2007). Climate Change. The Intergovernmental oanel on climate change mitigation of climate change. Cambridge, UK: Cambridge University Press.
- Jones, P.D. (2013). Green house effect and climate data. *Reference Mod. Earth Syst. Environ. Sci.*, 1-17.
- Kim, Y. (2015). Effect of thaw depth on fluxes of CO<sub>2</sub> dan CH<sub>4</sub> in manipulated Arctic coastal tundra of Barrow, Alaska. *Sci. Tot. Environ.*, **505**: 385-389.
- Kone, Y.J.M. and A.V. Borges (2008). Dissolved inorganic carbon dynamics in the waters surrounding forested mangroves of the Ca Mau Province (Vietnam). *Estuar: Coast. Shelf. Scie.*, 77: 409-421.
- Konnerup, D., J.M.B. Portela, C. Villamil and J.P. Parra (2014). Nitrous oxide and methane emissions from the restored mangrove ecosystem of the Ciénaga Grande de Santa Marta, Colombia. *Estuar. Coast. Shelf. Scie.*, **30:** 1-9.
- Kreuswiezer, J., J. Buchholz and H. Rennenberg (2003). Emission of methane and nitrous oxyde by Australian mangrove ecosystems. *Plant Biol.*, **5**: 423-431.
- Kristensen, E., S. Bouillon, T. Dittmar and C. March (2003). Organic carbon dynamics in mangrove ecosystems: a review. *Aquat. Bot.*, 89: 201-219.
- Kristensen, E., M.R. Flindt, S. Ulomi, A.V. Borges, G. Abril and S. Bouillon (2008). Emission of CO<sub>2</sub> and CH<sub>4</sub> to the atmosphere by sediments and open waters in two Tanzanian mangrove forests. *Mar. Ecol. Prog. Ser.*, **370**: 53-67.
- Lang, M., Z.C. Cai and S.X. Chang (2011). Effects of land use type and incubation temperature on greenhouse gas emissions from Chinese and Canadian Soils. J. Soil. Sed., 11: 15-24.
- Lekphet, S., S. Nitisoravut and S. Adsavakulchai (2005). Estimating methane emission from mangrove area in Ranong Province, Thailand. *Songklanakarin J. Sci. Technol.*, **27:** 153-163.
- Li, Y., S. Dong, S. Liu, H. Zhou, Q. Gao, G. Cao, X. Wang, X. Su, Y. Zhang, L. Tang and H. Zhao (2015). Seasonal changes of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O fluxes in different types of alpine grassland in the Qinghai-Tibetan Plateau of China. *Soil Biol. Biochem.*, **80**: 306-314.
- Lovelock, C.E., I.C. Feller, K.L. Mckee, B.M.J. Engelbrecht and M.C. Ball (2004). The effect of nutrient enrichment on growth, photosynthesis and hiudralic conductance of dwarf mangroves in Panama. *Funct. Ecol.*, **18**: 25-33.
- Martuti, N.K.T., D.L. Setyowati and S.B. Nugraha (2019).

Mangove Ecosystems: Diversity, Fitoremediation, Carbon Stocks, Function and their Management. Semarang, ID: LPPM – State University of Semarang.

- Mastepanov, M., C. Sigsggard, E.J. Dlugokencky, S. Houweling, L. Ström and M.P. Tamstorf (2008). Large tundra methane burst during onset of freezing. *Nat.*, 456: 628-631.
- Middelburg, J.J., J. Nieuwenhuize, N. Iversen, H. Høgh, H. de Wilde, W. Helder, R. Seifert and O. Christof (2002). Methane distribution in European tidal estuaries. *Biogeochem.*, 59: 95-119.
- Nirmal-Rajkumar, A., J. Barnes, R. Ramesh, R. Purvaja and R.C. Upstill-Goddard (2008). Methane and nitrous oxide ûuxes in the polluted Adyar River and estuary, SE India. *Mar. Pollut. Bull.*, 56: 2043-2051.
- Nobrega, G.N., T.O. Ferreira, M.S. Neto, H.M. Queiroz, A.G. Artur, E.D.S. Mendonca, E.D.O. Silva and X.L. Otero (2016). Edaphic factors controlling summer (rainy season) greenhouse gas emissions (CO<sub>2</sub> and CH<sub>4</sub>) from semiarid mangrove soils (NE-Brazil). *Sci. Tot. Environ.*, **542**: 685-693.
- Noor, Y.R., M. Khazali, I.N.N. Suryadiputra (2006). Guideline to Mangrove Introduction in Indonesia. Bogor, ID: Wetlands International – Indonesia Programme.
- Oertel, C., J. Matschullat, K. Zurba, F. Zimmermann and S. Erasmi (2016). Greenhouse gas emissions from soils- a review. *Chemie. der. Erde.*, **76:** 327-352.
- Ohwayo, R.O., V. Natugonza, L. Musinguzi, M. Olokotum and S. Naigaga (2016). Implications of climate variability and change for African lake ecosystems, fisheries productivity and livelihoods. J. Great Lakes Res., 42: 498-510.
- Oksanen, P.O., P. Kuhry and R.N. Alekseeva (2001). Development of the Rogovaya River peat plateau European Russian Arctic. *Holoc.*, **11**: 25-40.
- Ostrovsky, I. (2003). Methane bubbles in Lake Kinneret: quantification and temporal and spatial heterogeneity. *Limnol. Oceanogr.*, **48:** 1030-1036.
- Osudar, R., A. Matousu, M. Alawi, D. Wagner and I. Bussmann (2015). Environmental factors affecting methane distribution and bacterial methane oxidation in the German Bight (North Sea). *Estuar. Coast. Shelf. Sci.*, **160**: 10-21.
- Pathak, H. (1999). Emission of nitrous oxide from soil. Article Reviews, Curr. Sci., 77: 355-360.
- Purvaja, R. and R. Ramesh (2001). Natural and antrhropogenic methane emission from coastal wetlands of South India. *Environ. Manage.*, 27: 547-557.
- Rahman, D. Yanuarita and N. Nurdin (2014). Mangrove community structure in Muna Regency. *Torani.*, 24: 29-36.
- Rahman, F. Yulianda, I. Rusmana and Y. Wardiatno (2018). Fluxes of greenhouse gases CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O from mangrove soil in Tallo River, Makassar. *J. Trop. Biol.*, **18**: 149-158.
- Rahman, F. Yulianda, I. Rusmana and Y. Wardiatno (2019). Production ratio of seedlings and density status of mangrove ecosystem in coastal areas of Indonesia. *Advan. Environ. Biol.*, **13**: 13-20.
- Rosegrant, M.W., M.M. Dey, R. Valmonte-Santos and O.L. Chen (2016). Economic impacts of climate change and

climate change adaptation strategies in Vanuatu and Timor Leste. *Mar. Pol.*, **67:** 179-188.

- Shaffril, H.A.M., A.A. Samah and J.L. D'Silva (2017a). Adapting towards climate change impacts: Strategies for small-scale ûshermen in Malaysia. *Mar. Pol.*, 81: 196-201.
- Shaffril, H.A.M., A.A. Samah and J.L. D'Silva (2017b). Climate change: social adaptation strategies for fisherman. *Mar. Pol.*, 81: 256-261.
- Shawket, N., Y. Elmadhi, K.E. Kharrim and D. Belghyti (2019). Impacts of climate change on fish performance. *J. Entomol. Zoo. Stud.*, 7: 343-349.
- Singh, A.S., A. Zwickle, J.T. Bruskotter and R. Wilson (2017). The perceived psychological distance of climate change impacts and its inûuence on support for adaptation policy. *Environ. Sci. Pol.*, **73**: 93-99.
- Sotomayor, D., J.E. Corredor and J.M. Morell (1994). Methane flux from mangrove soils along the southwestern coast of Puerto Rico. *Estuar*, **17**: 140-147.
- Tam, N.Y.F. and Y.S. Wong (2000). Spatial variation of heavy metals in surface soils of Hong Kong mangrove swamps. *Environ. Pollut.*, **110**: 195-205.
- Trott, L.A. and D.M. Alongi (2000). The impact of shrimp pond effluent on water quality and phytoplankton biomass in a tropical mangrove estuary. *Mar. Pollut. Bul.*, **40**: 947-951.
- Tullberg, J., D.L. Antille, C. Bluett, J. Eberhard and C. Scheer (2018). Controlled traffic farming effects on soil emmissions of nitrous oxide and methane. *Soil. Till. Res.*, **176**: 18-25.
- Turunen, J., T. Tahvanainen, K. Tolonen and A. Pitkanen (2001). Carbon accumulation in West Siberian mires. Russia. *Glob. Biogeochem. Cyc.*, **15:** 761-769.
- Upstill-Goddar, R.C. and J. Barnes (2016). Methane emmisions from UK estuaries: Re-evaluating the estuarine source of tropospheric methane from Europe. *Mar. Chem.*, **180**: 14-23.
- Valmonte-Santos, R., M.W. Rosegrant and M.M. Dey (2016). Fisheries sector under climate change in the coral triangle countries of Paciûc Islands: Current status and policy issues. *Mar. Pol.*, 67: 148-155.
- Wang, S., C. Liu, K.M. Yeager, G. Wan, J. Lie and F. Tao (2009). The spatial distribution and emission of nitrous oxyde (N<sub>2</sub>O) in a large euthrophic lake in eastern China: anthropogenci effect. *Sci. Total Environ.*, **407**: 3330-3337.
- Wang, H., S. Zhou, X. Li, H. Liu, D. Chi and K. Xu (2016). The inûuence of climate change and human activities on ecosystem service value. *Ecol. Engin.*, 87: 224-239.
- Yunasfi (2006). Decomposition of mangrove litter leaf of Avicennia marina on various salinity level. [Dissertassion]. Bogor, ID: IPB University.
- Ye, Y., C. Lu, P. Yin, F. Tan and Y. Wong (2000). Diurnal charge of CH<sub>4</sub> fluxes from estuarine mangrove wetland. *Acta. Oceanol. Sin.*, **22**: 103-109.
- Zheng, X.H., M.X. Wang, Y.S. Wang, R.X. Shen, J. Gou, J. Li, J.S..Sin and L.T. Li (2000). Impacts of soil moisture on nitrous oxide emission from croplands: A case study on the rice-based agro-ecosystem in Southeast China. *Chemos-Glob. Change Sci.*, 2: 207-224.